



Policy analysis

A modelled global distribution of the kelp biome

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ABSTRACT

Kelp, seaweeds of the order Laminariales, are of ecological and conservation importance because they form undersea forest habitat for many varieties of fauna and flora including mammals, and commercial fish species. In the absence of a world map of the kelp biome, we predicted its potential distribution using geographic records and environment variables in a MaxEnt model. This estimated that the kelp biome occupied 1,469,900 km² and was present on 22% of the world's coastline. While average sea surface temperature was the most important environmental variable for the biome across all species, wave height, distance from the coast and minimum temperature were of most importance for individual species. This map can be used in planning where marine reserves should be best located, modelling the effects of climate change, and in estimating the blue (ocean) carbon storage. Current field observations should confirm the presence of kelp within the modelled biome, and if absent consider if human impacts, including climate change, are to blame.

1. Introduction

The kelp biome is comprised of over one hundred species of habitat-forming seaweeds of the order Laminariales many of which form forests on shallow rocky seabeds in temperate and subpolar seas, and a few deep cold tropical locations (Steneck et al., 2002; Graham et al., 2007; Krumhansl et al., 2016; Wernberg and Filbee-Dexter, 2019; Jayatilake and Costello, 2020c). The complexity of the three-dimensional structure of the kelp biome provides habitat for a diversity of species, including commercial fish (Teagle et al., 2017; Vásquez et al., 2014) and mammals of conservation importance (e.g., Markel and Shurin, 2015). Kelp forests are the dominant primary producers in cold-temperate rocky reef ecosystems (Krumhansl and Scheibling, 2012; Krumhansl et al., 2016) and amongst the most productive vegetation in the world (Mann, 1973). Not only does kelp have indirect benefits to society by virtue of its ecological importance, but some species provide food for people (Stévant et al., 2018; Peteiro and Freire, 2012). Kelp forests, and thus their associated fauna and flora, are threatened by harvesting, diseases, herbivory, competition from non-native species, storms, climate change, pollution, and the combined effects of these factors (Steneck et al., 2002; Wernberg et al., 2011; Krumhansl et al., 2016; Wernberg et al., 2019). Maps are thus useful to indicate the potential area that could be occupied by kelp species from local to global scales and facilitate conservation of kelp and

its associate faunal and floral communities.

Even though kelp does not have a root system to store carbon in sediments like in the seagrass and mangrove biomes, or a calcareous skeleton as in the zooxanthellate coral biome, the “blue” (ocean) carbon in kelp can be transformed to other food webs through herbivory and detrital pathways (Krumhansl and Scheibling, 2012; Alongi, 2018). Algal carbon sequestration occurs primarily through the burial of dead algae in sediments, and 82% of kelp productivity is estimated to become detritus (Krause-Jensen and Duarte, 2016; Krumhansl and Scheibling, 2012). However, the amount of carbon that is being stored by kelp alone and contributing to the carbon cycle of the ocean has not yet been quantified (Krause-Jensen and Duarte, 2016; Duarte, 2017). The global area of kelp occurrence is an important factor in this calculation.

To date, there is no existing map that can be used to calculate the global distribution of the kelp biome. A few hand-drawn maps have shown the distribution of the kelp genera *Macrocystis*, *Nereocystis*, *Laminaria*, *Lessonia* and *Ecklonia* (Raffaelli and Hawkins, 1999); and *Laminaria*, *Saccharina*, *Macrocystis*, *Nereocystis*, *Lessonia*, *Ecklonia* and *Eularia* (Wernberg et al., 2019). While useful to get an idea of the distribution of kelp, these maps cannot be used to calculate the biome area. Local-scale species distribution models have mapped the present distribution of kelp species and long term changes of kelp cover (Raybaud et al., 2013; Espriella et al., 2019), and locations of deep water tropical

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kelp refugia (Graham et al., 2007). However, none of these studies have developed a polygon layer of the global distribution of kelp biome that can be used in geographical information systems (GIS).

The current study fills the above mentioned research gap by modelling the global distribution of the kelp biome using field records and environmental variables to provide a world map. The combination of observed locations of species and knowledge of their environmental preferences enables the use of species distribution models to more comprehensively map this biome. The availability of such a global map will illustrate the importance of the kelp biome in global biodiversity, and enable improved estimates of global primary production, blue carbon budget, and deforestation rates. Further, this map will be useful for mapping the distribution of kelp associated fauna.

2. Methods

2.1. Species occurrence data

The common term kelp typically refers to the order Laminariales. However, sometimes some large brown algae in the order Fucales, such as species of the genus *Durvillaea*, are included (Dayton, 1985; Fraser et al., 2009; Wernberg and Filbee-Dexter, 2019). The present study was limited to the order Laminariales, which has 59 genera and 147 species (Guiry and Guiry, 2018). Kelp distribution data were extracted from the Global Biodiversity Information Facility (GBIF, 2017) and the Ocean Biogeographic Information System (OBIS, 2017). Initially, we downloaded 109,824 occurrence records for the order Laminariales in 145 datasets from GBIF, and 47,695 records in 99 datasets from OBIS (SM Appendix 3). Prior to analysis, taxonomic names were reconciled with AlgaeBase and the World Register of Marine Species (WoRMS) (Guiry and Guiry, 2018; Guiry and Guiry, 2020; Horton et al., 2020). Since our analysis, *Chorda filum* has been removed from the Laminariales into the new brown algae order, Chordales but is still included in our dataset (Starko et al., 2019; Guiry and Guiry, 2020). Data were available for 70 species belonging to 5 families (Table 1). We limited the model training dataset to data collected from 1900 to 2017. Of the data used, 86% were sampled between 1970 and 2017. Records of fossil specimens, ambiguous locations according to comments in the dataset (e.g., drift material), and records with coordinate uncertainty >10 km and falling on land, were removed. Multiple records for a species at the same location (i.e., identical latitude and longitude), including duplicates from both databases, were reduced to one record for that location (Fig. 1). We excluded 112,890 (72%) data points from the analysis as being questionable. The dataset used in this study can be downloaded from Figshare (Jayatilake and Costello, 2020a). In addition, we have downloaded and mapped more recently published data from the GBIF (temporal scale from 2018 to 2020; spatial scale 90 N to 90 S and 180 E to 180 W) to assess the accuracy of the predicted model. This was prepared following the same procedure to use as the accuracy assessment dataset and is also available on Figshare (Jayatilake and Costello, 2020b).

2.2. Environmental data

We obtained the environmental data from Global Marine Environment Datasets (GMED) (Basher et al., 2014b; Basher and Costello, 2019). These data layers represent annual averages calculated over decades and thus indicate long-term characteristics of the environment (Basher and Costello, 2019) (Table S1). GMED environmental layers have a 5 arc-minute resolution which is approximately 9.2 km at the equator (Table S1). For this study, a finer spatial resolution was needed to get more accurate distributions. Therefore, we re-interpolated the GMED data to 30 arc-second resolution which is approximately 1 km at the equator. All the interpolated raster layers were cropped to a 0 to 1000 m depth layer to reduce the computational time. The deepest at which any species of kelp has been found to be living is 90 m for *Laminaria*

rodriguezii in the Adriatic Sea (Žuljević et al., 2016). Deeper records may be due to the sinking of kelp from shallow depths following detachment due to storms. Graham et al. (2007) predicted that the maximum depth for any kelp species would be 236 m in tropical deep waters. Thus, the present study extended well beyond the deepest likely range of kelp.

Previous studies focused on the influence of single abiotic variables such as temperature, wave exposure, water motion, salinity, light availability, dissolved oxygen, pH, nitrate and phosphate on the natural distribution of one or a few kelp species (Dayton, 1985; Graham et al., 1997; Gerard, 1997; Hurd, 2000; Gaylord et al., 2002; Steneck et al., 2002; Wernberg and Thomsen, 2005; Smale et al., 2013; Žuljević et al., 2016; Wernberg et al., 2019). Here we selected almost all the abiotic variables from previous studies to study which variables most correlated with the distribution of kelp, noting that many variables are also correlated with each other (Table S2). A preliminary MaxEnt model was run with 19 environmental variables (mean sea bottom temperature, calcite concentration, depth, diffuse attenuation coefficient, dissolved oxygen concentration, distance from the land, nitrate concentration, pH, phosphate concentration, photosynthetically active radiation, average sea surface temperature, maximum sea surface temperature, minimum sea surface temperature, range of sea surface temperature, salinity, silicate concentration, slope, surface current, and wave height) for each species, genus, family and all kelp (Table 2). However, the variables calcite, minimum and range of sea surface temperature, mean sea bottom temperature, surface current, and silicate had <0.5% contribution to the models and were thus excluded from the kelp biome model.

2.3. Modelling

The Maximum Entropy (MaxEnt) modelling software has been widely applied for marine species distribution modelling with presence-only data (e.g., Tittensor et al., 2009; Verbruggen et al., 2009; Yesson et al., 2012; Basher et al., 2014a; Saeedi et al., 2016; Jayatilake and Costello, 2018; Martínez et al., 2018). We used MaxEnt version 3.3.3 k (Phillips et al., 2006; Phillips and Dudík, 2008) to generate the kelp distribution model. In the current study, the model had 10 replicate runs with cross-validation, a maximum number of background points 10,000, and maximum iterations 1000. The 'remove duplicate presence records' option was activated to keep one observation point per 30 arc sec grid cell. Separate MaxEnt models for species, genera, families and the order were created, and the environmental variables contributing most to the models tabulated. However, 25 species had insufficient occurrence records to develop individual MaxEnt models to predict their distributions.

By applying the model to all species together we were able to include records of even rare species. We previously found that this approach provides a more accurate model of a marine biome distribution (Jayatilake and Costello, 2018). While our model used 13 potentially related variables, we also generated results (see Model evaluation below) to determine the contribution of each variable individually.

2.4. Model evaluation

The accuracy of the MaxEnt model was evaluated using the receiver operating characteristic (ROC) curve and AUC (area under the ROC) (Peterson et al., 2008; Peterson et al., 2011). The ROC curve and AUC evaluate how well a species distribution model fits true presence and absence data (Elith et al., 2006; Elith et al., 2011). It is a graphical interpretation of the omission and commission rates. The omission rate is defined as the proportion of known occurrence records which are not predicted as presence. The proportion of known presence records predicted as present in the model is known as its sensitivity (1 - omission rate) (Phillips et al., 2004). Theoretically, the commission rate is the proportion of absences predicted as presence. The commission rate is defined as 1 - specificity, where specificity is the proportion of absences correctly predicted (Phillips et al., 2004). In a study where only presence

Table 1

List of laminarian kelp species and species names used in this study to model the global distribution of the kelp biome.

Agaraceae
<i>Agarum clathratum</i> Dumortier, 1822
<i>Agarum turneri</i> Postels & Ruprecht, 1840
<i>Neogagarum fimbriatum</i> (Harvey) H.Kawai & T.Hanyuda, 2017
<i>Costaria costata</i> (C. Agardh) De A. Saunders, 1895
<i>Dictyoneurum californicum</i> Ruprecht, 1852
<i>Dictyoneurum reticulatum</i> (De A.Saunders) P.C.Silva, 2008
<i>Thalassiophyllum clathrus</i> (S. G. Gmelin) Postels & Ruprecht, 184
Alariaceae
<i>Alaria angusta</i> Kjellman, 1889
<i>Alaria crassifolia</i> Kjellman, 1885
<i>Alaria crispa</i> Kjellman, 1889
<i>Alaria esculenta</i> (Linnaeus) Greville, 1830
<i>Alaria marginata</i> Postels & Ruprecht, 1840
<i>Alaria praelonga</i> Kjellman, 1889
<i>Alaria pylaïi</i> (Bory de Saint-Vincent) Greville, 1830
<i>Eualaria fistulosa</i> (Postels & Ruprecht) M. J. Wynne, 2009
<i>Lessoniopsis littoralis</i> (Farlow & Setchell ex Tilden) Reinke, 1903
<i>Pleurophycus gardneri</i> Setchell & Saunders ex Tilden, 1900
<i>Pterygophora californica</i> Ruprecht, 1852
<i>Undaria pinnatifida</i> (Harvey) Suringar, 1873
Chordaceae
<i>Chorda filum</i> (Linnaeus) Stackhouse, 1797 ^b
Laminariaceae
<i>Cymathære triplicata</i> (Postels & Ruprecht) J.Agardh, 1868
<i>Hedophyllum bongardianum</i> (Postels & Ruprecht) Yendo, 1914
<i>Hedophyllum dentigerum</i> (Kjellman) Starko, S.C.Lindstrom & Martone, 2019
<i>Hedophyllum sessile</i> (C.Agardh) Setchell, 1901
<i>Laminaria abyssalis</i> A.B.Joly & E.C.Oliveira, 1967
<i>Laminaria brasiliensis</i> A.B.Joly & E.C.Oliveira, 1967 ^a
<i>Laminaria digitata</i> (Hudson) J.V.Lamouroux, 1813
<i>Laminaria ephemera</i> Setchell, 1901
<i>Laminaria farlowii</i> Setchell, 1893
<i>Laminaria hyperborea</i> (Gunnerus) Foslie, 1884
<i>Laminaria longipes</i> Bory de Saint-Vincent, 1826
<i>Laminaria ochroleuca</i> Bachelot de la Pylaie, 1824
<i>Laminaria pallida</i> Greville, 1848
<i>Laminaria rodriguezii</i> Bornet, 1888
<i>Laminaria setchellii</i> P.C.Silva, 1957
<i>Laminaria sinclairii</i> (Harvey ex J.D.Hooker & Harvey) Farlow, Anderson & Eaton, 1878
<i>Laminaria solidungula</i> J.Agardh, 1868
<i>Laminaria yezoensis</i> Miyabe, 1902
<i>Macrocystis pyrifera</i> (Linnaeus) C. Agardh, 1820
<i>Nereocystis luetkeana</i> (K. Mertens) Postels & Ruprecht, 1840
<i>Pelagophycus porra</i> (Léman) Setchell, 1908
<i>Postelsia palmaeformis</i> Ruprecht, 1852
<i>Saccharina angustata</i> (Kjellman) C.E.Lane, C.Mayes, Druehl & G.W.Saunders, 2006
<i>Saccharina cichorioides</i> (Miyabe) C.E.Lane, C.Mayes, Druehl & G.W.Saunders, 2006
<i>Saccharina complanata</i> (Setchell & N.L.Gardner) Gabrielson, Lindstrom & O'Kelly, 2012
<i>Saccharina japonica</i> (J.E. Areschoug) C.E.Lane, C.Mayes, Druehl & G.W.Saunders, 2006
<i>Saccharina latissima</i> (Linnaeus) C.E.Lane, C.Mayes, Druehl & G.W.Saunders, 2006
<i>Saccharina longicurvis</i> (Bachelot de la Pylaie) Kuntze, 1891
<i>Saccharina nigripes</i> (J.Agardh) Lontin & G.W.Saunders, 2015
<i>Streptophyllopsis kuroshioense</i> (Segawa) Kajimura, 1981
Lessoniaceae
<i>Ecklonia biruncinata</i> (Bory) Papenfuss, 1944
<i>Ecklonia brevipes</i> J. Agardh, 1877
<i>Ecklonia cava</i> Kjellman, 1885
<i>Ecklonia fastigiata</i> (Endlicher & Diesing) Papenfuss, 1940
<i>Ecklonia kurome</i> Okamura, 1927
<i>Ecklonia maxima</i> (Osbeck) Papenfuss, 1940
<i>Ecklonia muratii</i> Feldmann
<i>Ecklonia radiata</i> (C. Agardh) J. Agardh, 1848
<i>Ecklonia richardiana</i> J. Agardh, 1848
<i>Ecklonia stolonifera</i> Okamura, 1913
<i>Egregia menziesii</i> (Turner) Areschoug, 1876
<i>Eisenia cokeri</i> M. Howe, 1914

(continued on next page)

Table 1 (continued)

Lessoniaceae
<i>Lessonia adamsiae</i> C. H. Hay, 1987
<i>Lessonia brevifolia</i> J. Agardh, 1894
<i>Lessonia corrugata</i> Lucas, 1931
<i>Lessonia flavicans</i> Bory de Saint-Vincent, 1825
<i>Lessonia nigrescens</i> Bory de Saint-Vincent, 1826
<i>Lessonia tholiformis</i> C. H. Hay, 1989
<i>Lessonia trabeculata</i> Villouta & Santelices, 1986
<i>Lessonia variegata</i> J. Agardh, 1878

^a Marins et al. (2012) considered *L. brasiliensis* a synonym of *L. abyssalis*.

^b Chorda filum has since moved to another Order (Starko et al., 2019).

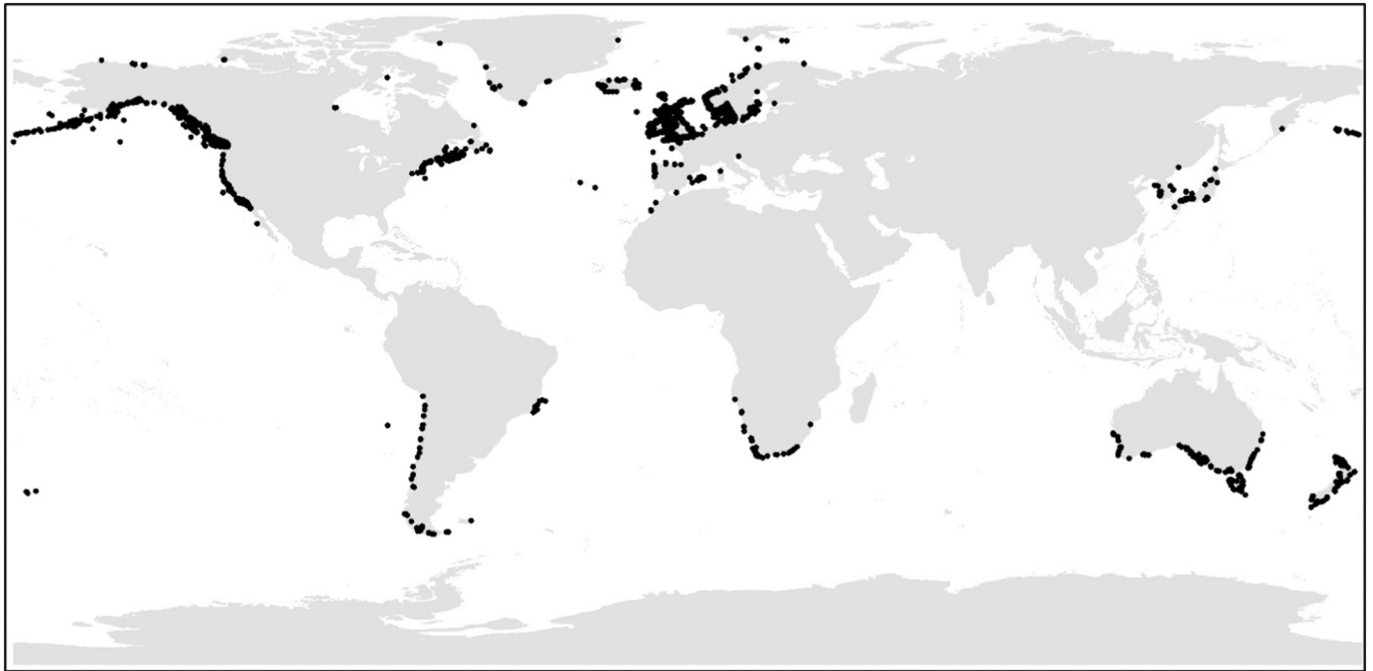


Fig. 1. The distribution of laminarian kelp observations used in this study.

Table 2

The environmental variables used in the Maxent models to predict the geographic distribution of kelp species of the order Laminariales. Columns indicate the range of each variable in the data used, the range kelp occurred most frequently in, and statistics on the relative contribution of each variable used to predicted the distribution of kelp. SST = sea surface temperature. PSS = practical salinity scale units.

Abiotic variable	Unit	Variable range	Most suitable range	Percent contribution
Average SST	°C	0–35	5–25	47.7
Land distance	km	0–20	0–1	23.7
Maximum SST	°C	0–35	7–27	15.9
Wave height	m	0–8	1–8	9.5
Dissolved oxygen	ml l ⁻¹	1–10	5–7	1.2
Depth	m	0–1000	0–100	0.7
Nitrate	µmol l ⁻¹	0–30	5–14	0.3
pH		6.6–8.6	7.9–8.3	0.3
Photosynthetically active radiation	Einstein/m ² /day	0–55	25–40	0.3
Phosphate	µmol l ⁻¹	0–2.5	0.1–1.0	0.1
Salinity	PSS	0–45	0–5 and 30–37	0.1
Slope	degree	0–14	0–3	0.1
Diffuse attenuation coefficient	m ⁻¹	0–65	0–25	0

data are available (i.e., no true absence data), MaxEnt selects random background points as pseudo-absences instead of true absence records (Phillips and Dudík, 2008). Here it assumes that, all the grid cells without occurrence localities could be pseudo absences even if they have suitable environmental conditions (Phillips et al., 2004; Phillips et al., 2006). The current study area was ten times deeper than the average depth of kelp. This greatly increased the likelihood that the location that would be selected as a pseudo-absence would be a true absence.

MaxEnt has a high predictive performance with presence-only data (Elith et al., 2006). AUC is an indicator of the predictive power of a probabilistic model and ranges from 0 to 1, where the highest ranking is 1 (Phillips et al., 2004; Phillips and Dudík, 2008; Peterson et al., 2008). The MaxEnt model indicates which variables best explained the distribution of the species using analyses of percent contribution, response curves, and a jack-knife test. While the MaxEnt model is being run the percentage contribution of each variable to the model is calculated. This gives a heuristic estimate of the relative contribution of the environmental variable to the MaxEnt model (Phillips, 2017). The jack-knife test creates three plots to show how each variable has contributed to model training, model testing and the AUC. By evaluating the overall results of each jack-knife plot we can predict which of the variables mattered the most in determining an environmental or geographic distribution.

The post image processing of the MaxEnt modelled map was done using the ArcGIS version 10.5.1. The MaxEnt probability values above

0.45 gave the most visibly similar geographic coverage to the field observation records. Those areas were considered to define the global distribution of the kelp biome. The accuracy of the classified map was cross-checked with another separate dataset of the order Laminariales downloaded from the GBIF (from 2018 to 2020) (Jayatilake and Costello, 2020b). The MaxEnt probability values for occurrence records were extracted using the ArcGIS tool “raster value to point”. The percentage of the occurrence records plotted within the predicted area were calculated using the MaxEnt probability values of these new occurrence records. MaxEnt probability values of these records were given in the last column of the table available in Jayatilake and Costello (2020b). The original abiotic layers were on the WGS84 geographical coordinate system. Thus, the initial MaxEnt modelled map used WGS84 geographical coordinate system which has larger grid cells at lower latitudes. We converted the MaxEnt modelled map to cylindrical equal-area projection (all grid cells have the same area) using the ArcGIS projection tool to calculate the true area of distribution and the coastline length covered by the kelp biome.

3. Results

At the order level, the laminarian kelp biome data contained 44,265 occurrence records distributed mainly in temperate and sub-Polar Regions. There were no occurrence records from the tropics and Antarctica (Fig. 1).

The modelled kelp biome map closely matched the distribution of reported occurrence records (Figs. 1 and 2). The high AUC indicated that the model had a probability of 0.771 to discriminate predicted presence records over the pseudo-absence records. In the validation dataset 86% (2626 out of 3054 occurrence records from 2018 to 2020) were plotted within the area predicted by a MaxEnt probability value of ≥ 0.45 and the remaining area at a spatial resolution of 30 arcsec (Fig. 3).

The biome covered 1,469,900 km² and 22% of the world's coastline.

The modelled map predicted the distribution of kelp mainly in the temperate, sub-Polar and the Arctic Ocean. The model predicted locations suitable for kelp which lacked georeferenced records in GBIF and OBIS, namely: the Atlantic coast of Argentina; Hokkaido Island, Japan; Shandong Peninsula, China; and Svalbard Island in the Arctic Ocean. However, the model did not predict laminarian kelp forests in the tropics and Antarctica.

The annual average sea surface temperatures (SST), distance from land, the maximum sea surface temperature and wave height, were the topmost variables contributing to the MaxEnt model (Table 2). The environmental variable with the highest gain when used in isolation was the annual average SST. Thus, the average SST had the most useful information by itself. Distance from the land was the third, wave height was the fourth, and dissolved oxygen the fifth most important variable for predicting the distribution of kelp (Table 2).

The probability of occurrence of kelp decreased with depth from 0 to 100 m, and no kelp occurred deeper than 250 m (Fig. 4). Most kelp occurred within 1 km of land. Kelp largely occurred with a maximum SST from 7 °C to 27 °C, and average from 5 °C to 25 °C. Kelp never occurred above an annual maximum of 30 °C and an annual average of 27 °C. The probability of kelp occurrence increased with wave height up to 7 m. Although no kelp was predicted above a salinity of 37.5, there were peaks of occurrence at 5 and 35 PSS (Fig. 4). The low salinity peak was due to the presence of *Chorda filum* in the Baltic Sea, parts of which have low salinity. Note that this species has recently been moved into a new order outside Laminariales (Starko et al., 2019).

The results from the training gain, test gain, and AUC jack-knife test plots showed a similar pattern of the contribution of each variable for the model. The average and maximum SST gave higher regularized training, test gain, and AUC compared to other variables (Fig. 5). The next most important variables were dissolved oxygen, wave height and the distance from land. If MaxEnt used only slope, salinity and pH, there was almost no gain in all three plots (Fig. 5). Thus, these three variables

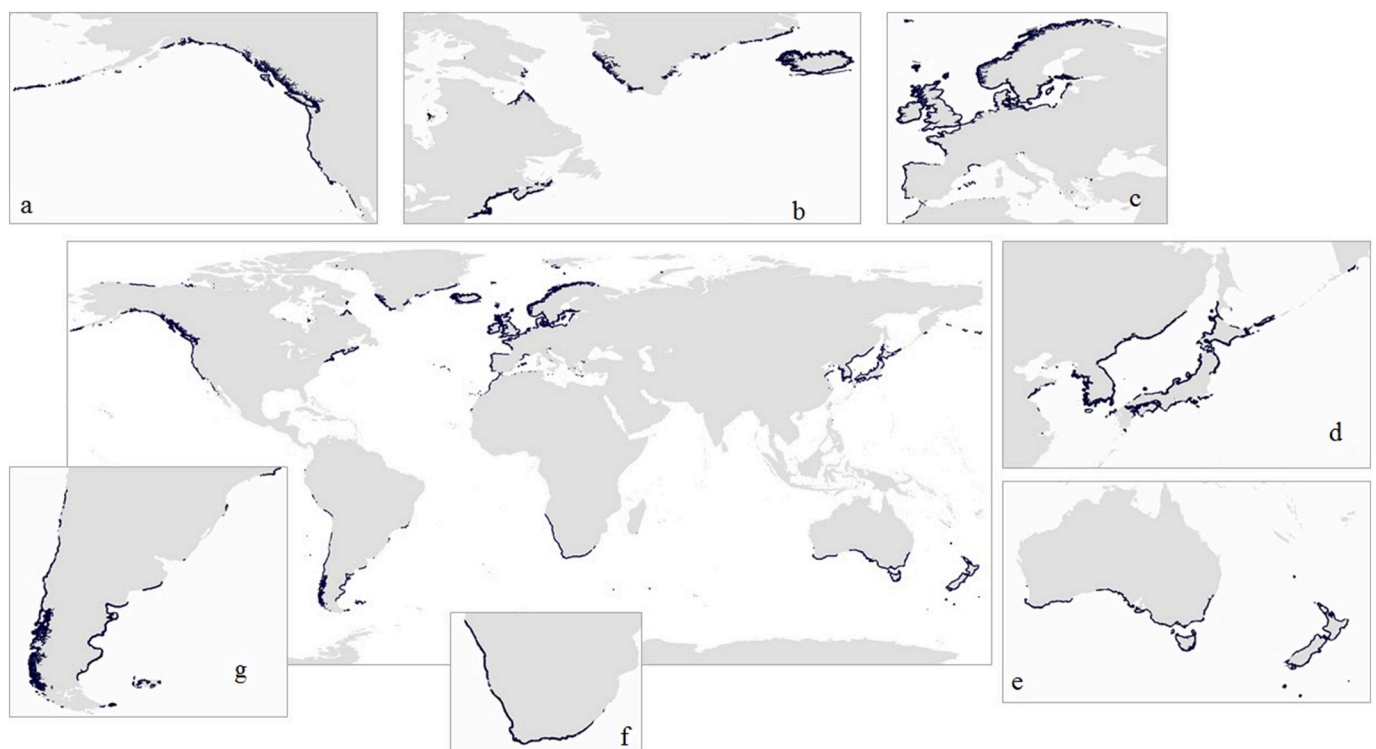


Fig. 2. The predicted environmental range for the order Laminariales. The dark blue colour indicates the MaxEnt probability of distribution. (a) west coast of North America, (b) north-west Atlantic including Greenland and Iceland, (c) Europe, (d) north-west Pacific including parts of Japan, China, Russia, and Korea, (e) New Zealand and southern Australia, (f) southern Africa, (g) southern South America. (For interpretation of the references to colour in this figure legend, the reader is referred to the web version of this article.)

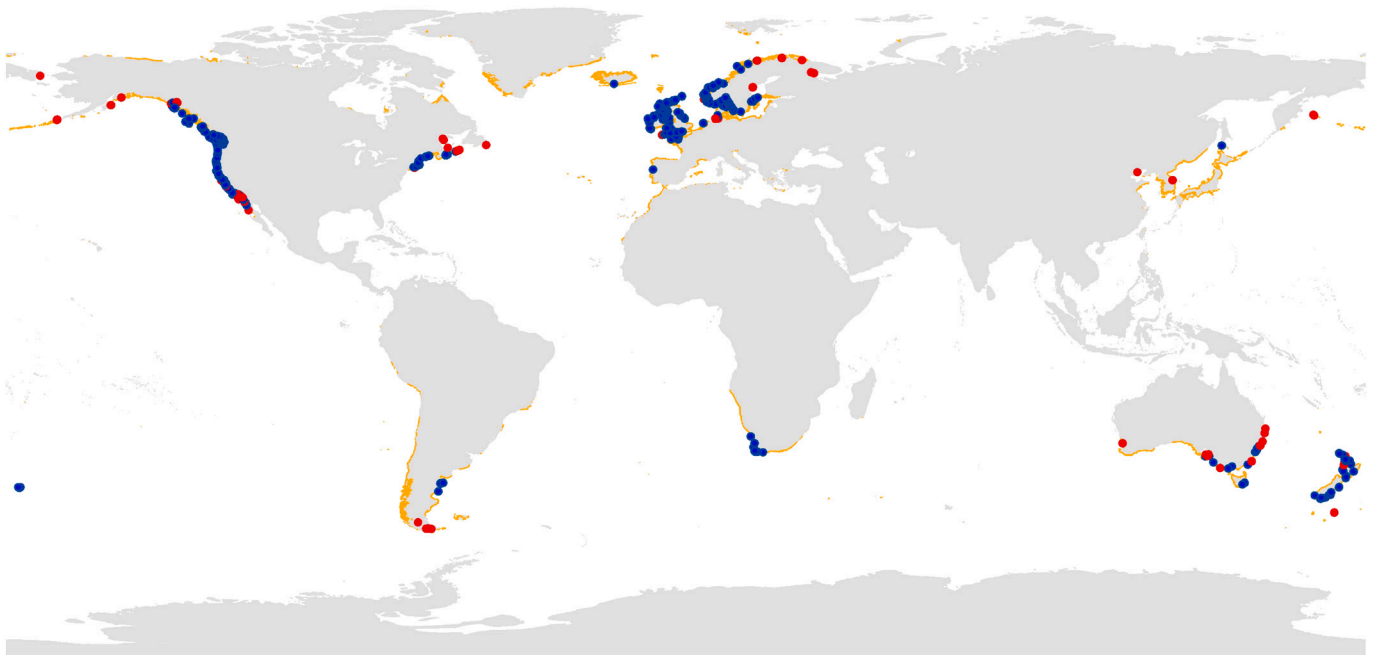


Fig. 3. The comparison of the predicted laminarian kelp biome (yellow) to the occurrence of additional data published in since 2018 that were not used in modelling the biome; 86% of these new records occurred within the biome. The blue colour points indicate the occurrence records have ≥ 0.45 maxent probability value (plotted within the predicted area) and the red colour points show the occurrence records have < 0.45 MaxEnt probability values (plotted out of the predicted area). (For interpretation of the references to colour in this figure legend, the reader is referred to the web version of this article.)

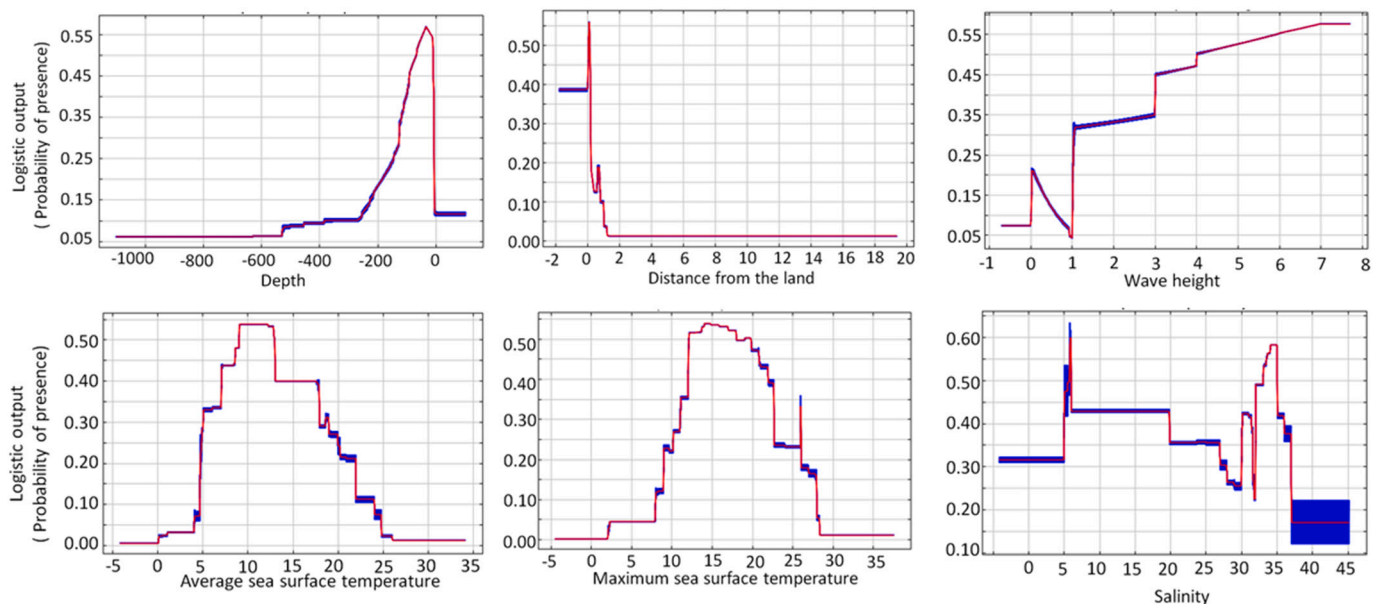


Fig. 4. The response of kelp to depth, distance from land, wave height, average sea surface temperature, maximum sea surface temperature, and salinity.

were not meaningful for predicting the global distribution of kelp.

Of the 70 species used in this study for which geographic coordinates were available, 46 species had insufficient georeferenced records to be successfully modelled. Wave height was the topmost for 23, and one of the top three most important environmental variables for 35 of these species in the MaxEnt models (Table 3). Distance from the land was the next most important variable, being amongst the top three variables for 21 species, followed by minimum SST (14 species). Wave height and land distance, followed by minimum and average SST, were also the most important variables at the genus level (Table 4). Wave height, average SST and minimum SST were the most important variables at the

family level (Table 5).

4. Discussion

In this study, we provide the first global distribution map of laminarian kelp as a polygon layer that can be used in geographical information systems (GIS). This polygon layer has a more complete geographical distribution of the kelp biome than the published range maps of kelp species. The map had a very similar distribution to the observed field records (Fig. 1), and as reported in the literature (Steneck et al., 2002; Wernberg et al., 2019; Wernberg and Filbee-Dexter, 2019).

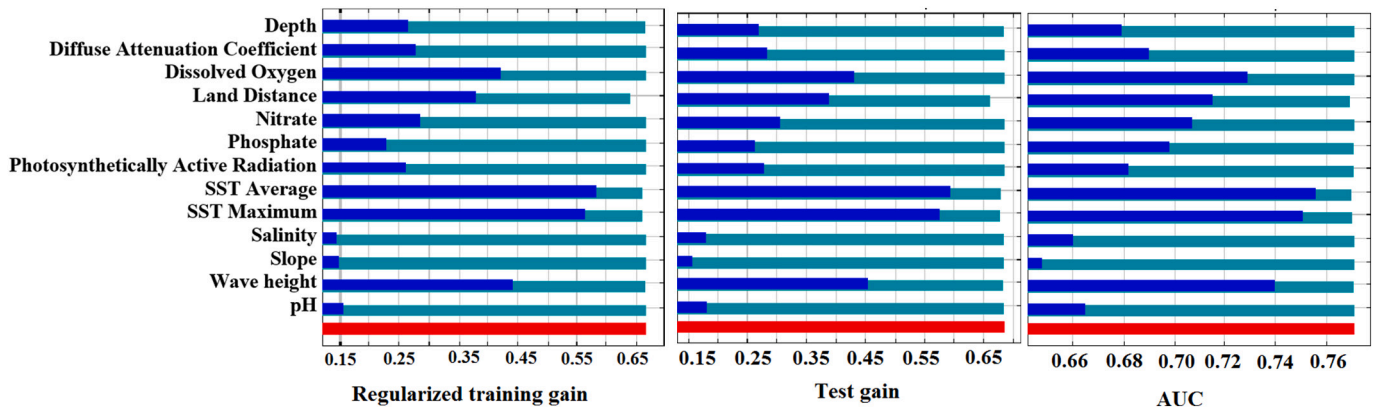


Fig. 5. The results of the jack-knife tests of variable importance: (a) training gain; (b) test gain, (c) AUC. Jack-knife results were calculated without the variable (green), with only variable (blue) and with all variables (red). (For interpretation of the references to colour in this figure legend, the reader is referred to the web version of this article.)

Table 3

Estimates of relative contributions of the environmental variables to the MaxEnt model of the laminarian kelp species. SST = Sea Surface Temperature, SBT = Sea Bottom Temperature, DAC = Diffuse Attenuation Coefficient, PAR = Photosynthetically Active Radiation.

Species	1st	2nd	3rd
<i>Agarum clathratum</i>	Land distance	Wave height	Maximum SST
<i>Agarum turneri</i>	Depth	Phosphate	Wave height
<i>Costaria costata</i>	Wave height	Land distance	Salinity
<i>Dictyoneuropsis reticulata</i>	Wave height	Land distance	Minimum SST
<i>Alaria crista</i>	Wave height	Phosphate	Land distance
<i>Alaria esculenta</i>	Wave height	Mean SBT	Nitrate
<i>Alaria marginata</i>	Wave height	Land distance	Minimum SST
<i>Alaria praelonga</i>	Wave height	Land distance	Phosphate
<i>Eualaria fistulosa</i>	Wave height	Land distance	Phosphate
<i>Lessoniopsis littoralis</i>	Wave height	Land distance	Salinity
<i>Pleurophycus gardneri</i>	Wave height	Land distance	Salinity
<i>Pterygophora californica</i>	Wave height	Minimum SST	Land distance
<i>Undaria pinnatifida</i>	Land distance	Wave height	Minimum SST
<i>Chorda filum</i>	Mean SBT	Wave height	Land distance
<i>Laminaria abyssalis</i>	Minimum SST	Wave height	Nitrate
<i>Laminaria brasiliensis</i>	Minimum SST	Dissolved oxygen	Wave height
<i>Laminaria digitata</i>	Average SST	Wave height	Land distance
<i>Laminaria ephemera</i>	Wave height	Land distance	pH
<i>Laminaria hyperborea</i>	Wave height	Average SST	Minimum SST
<i>Laminaria ochroleuca</i>	Average SST	Nitrate	Phosphate
<i>Laminaria pallida</i>	PAR	Wave height	Maximum SST
<i>Laminaria rodriguezii</i>	Average SST	Minimum SST	Nitrate
<i>Laminaria setchellii</i>	Wave height	Land distance	Maximum SST
<i>Laminaria sinclairii</i>	Wave height	Maximum SST	Land distance
<i>Laminaria solidungula</i>	Depth	pH	Maximum SST
<i>Laminaria yezoensis</i>	Wave height	Phosphate	Land distance
<i>Macrocystis pyrifera</i>	Wave height	Land distance	Minimum SST
<i>Nereocystis luetkeana</i>	Wave height	Land distance	Minimum SST
<i>Pelagophycus porra</i>	Nitrate	Wave height	Slope
<i>Postelsia palmaeformis</i>	DAC	Wave height	Minimum SST
<i>Saccharina dentigera</i>	Wave height	Land distance	Nitrate
<i>Saccharina latissima</i>	Wave height	Average SST	Land distance
<i>Saccharina sessilis</i>	Wave height	Land distance	DAC
<i>Ecklonia cava</i>	Maximum SST	Phosphate	Wave height
<i>Ecklonia kurome</i>	Maximum SST	Depth	PAR
<i>Ecklonia maxima</i>	Wave height	PAR	Phosphate
<i>Ecklonia radiata</i>	Salinity	Wave height	Land distance
<i>Egregia menziesii</i>	Wave height	Land distance	Minimum SST
<i>Lessonia corrugate</i>	Wave height	Minimum SST	Land distance
<i>Lessonia flavicans</i>	Land distance	Phosphate	Dissolved oxygen
<i>Lessonia variegata</i>	Minimum SST	Land distance	Nitrate

Table 4

Estimates of relative contributions of the environmental variables to the MaxEnt model of the laminarian kelp genera. SST = Sea Surface Temperature.

Genus	1st most contribution	2nd most contribution	3rd most contribution
<i>Agarum</i>	Wave height	Land distance	Salinity
<i>Alaria</i>	Average SST	Wave height	Land distance
<i>Chorda</i>	Average SST	Wave height	Land distance
<i>Costaria</i>	Wave height	Land distance	Salinity
<i>Dictyoneuropsis</i>	Wave height	Land distance	Minimum SST
<i>Ecklonia</i>	Minimum SST	Land distance	pH
<i>Egregia</i>	Wave height	Land distance	Diffuse attenuation coefficient
<i>Laminaria</i>	Average SST	Wave height	Land distance
<i>Lessonia</i>	Land distance	Minimum SST	Wave height
<i>Macrocystis</i>	Wave height	Land distance	Minimum SST
<i>Nereocystis</i>	Wave height	Land distance	Nitrate
<i>Postelsia</i>	Diffuse attenuation coefficient	Wave height	Minimum SST
<i>Pterygophora</i>	Wave height	Minimum SST	Diffuse attenuation coefficient
<i>Saccharina</i>	Wave height	Average SST	Land distance
<i>Thalassiophyllum</i>	Average SST	Phosphate	Minimum SST
<i>Undaria</i>	Land distance	Wave height	Average SST

Table 5

Estimates of the relative contributions of the environmental variables to the MaxEnt model of the laminarian kelp families. SST = Sea Surface Temperature.

Family	1st most contribution	2nd most contribution	3rd most contribution
Agaraceae	Wave height	Land distance	Average SST
Alariaceae	Wave height	Average SST	Land distance
Chordaceae	Average SST	Land distance	Wave height
Laminariaceae	Average SST	Wave height	Land distance
Lessoniaceae	Minimum SST	Land distance	Wave height

Our model predicted that kelp was limited to latitudes 25° to 70° in the northern, and 25° to 55° in the southern, hemispheres. Of 3000 newly recorded occurrences from 2018 to 2020 86% were plotted on the biome and 14% were plotted nearby (Figs. 2 and 3). Such variability is to be expected considering both the spatial resolution of the coastline and environmental data, and variance in reporting latitude and longitude coordinates.

Kelp occupied 1,469,900 km² and 22% of the world's coastline. Previous studies estimated that 25% of the world's coastline was covered by kelp forests (Filbee-Dexter and Wernberg, 2018; Wernberg et al., 2019). Thus, the kelp biome is the second most widely distributed

marine biome, following seagrass with 1,646,788 km² (Jayatilake and Costello, 2018). Following the usage in terrestrial ecology, equivalent marine biomes are large areas characterised by plants of similar life-form that provide enduring three-dimensional habitat for other species (Woodward et al., 2004; Costello et al., 2020). The other marine biomes have ten times less area than kelp, namely zooxanthellate coral with 151,390 km² (UNEP-WCMC et al., 2018), and mangroves with 136,850 km² (Giri et al., 2011).

Some of the locations predicted to contain kelp in our map, but without occurrence records in GBIF and OBIS, were reported to have kelp forests in the literature. *Macrocystis pyrifera* and *Undaria pinnatifida* have been recorded in the Gulf of Nuevo, along the coast of Argentina from Puerto Deseado (Santa Cruz province) to Mar del Plata (Buenos Aires province) (Raffo et al., 2009; Pereyra et al., 2017; Paula et al., 2018). *Laminaria japonica* and *Saccharina japonica* occur along the coast of Shandong Peninsula, China (Wu et al., 2016; Shao et al., 2019). *Laminaria japonica*, *L. religiosa* and *U. pinnatifida* occur around Hokkaido Island, Japan (Matsunaga et al., 1999). *Alaria esculenta*, *Laminaria digitata*, and *Saccharina latissimi* occur in Hornsund, and *L. digitata* in Kongsfjorden, Svalbard (Włodarska-Kowalczyk et al., 2009; Bartsch et al., 2016). The current map did not predict any suitable locations in Antarctica and no laminarian kelp have been reported there (Moe and Silva, 1977; Quartino and Boraso de Zaiuso, 2008; Wernberg et al., 2019). This suggests that the absence of laminarian kelp in these regions is primarily due to environmental unsuitability. Thus, our map appears to be an accurate representation of the kelp biome on a global scale. However, the tropical deep-water kelp distributions were not predicted in this model due to a lack of occurrence records from the tropics and because the mean sea bottom temperature variable was excluded from the analysis due to its poor contribution to the model. These communities should be modelled separately with deep water variables.

As a photosynthetic plant, kelp is limited to the photic zone. In this study, the MaxEnt probability of the presence of kelp was high between 0 and 100 m depth, and it was limited to 1 km from the land (Fig. 4). However, it is likely that there may be offshore rocky reefs, such as the tops of seamounts, where kelp may occur but were not detected due to the spatial resolution of our data (Parker and Tunnicliffe, 1994; Bo et al., 2011). Kelp forests always occur on hard substrata such as rocky seabeds (Teagle et al., 2017; Wernberg et al., 2019). The present study could not include seabed substrata because a global layer is not available. Nevertheless, the accuracy of the map suggests that sufficient rocky substrata exist for all regions where temperature and light are suitable. However, more detailed regional maps of kelp distribution would benefit from including seabed substratum within the present biome map.

Kelp had an increased probability of presence with increasing wave height from 1 m to 7 m, with a low probability of occurrence in areas without wave action (Fig. 4). Most kelp species prefer turbulent water (Hurd, 2000; Wernberg et al., 2019). For the individual species, genera and families, wave height was generally the most important factor in influencing their distribution. Species such as *Laminaria hyperborea* and *Alaria esculenta* are more common on wave exposed coasts (Frid and Kitching, 1988; Norton, 1992; Kraan et al., 2000; Pedersen et al., 2012).

We confirmed that the annual average sea surface temperature is the most significant factor limiting the distribution of the kelp biome, as suggested by others (e.g., Lüning, 1990; Muller et al., 2009). Kelp occurred in average sea surface temperatures from 5 °C to 25 °C and was rare above 27 °C (Table 2). Thus, if sea surface temperature increases beyond these temperatures, such as due to global warming, it will alter the kelp distribution (Martínez et al., 2018; Assis et al., 2016). Indeed, the range of Australian temperate kelp forests has contracted after ocean warming and extreme heat waves (Wernberg et al., 2012; Wernberg et al., 2016). *Macrocystis pyrifera* forests in Australia have been predicted to disappear if the predicted high sea surface temperatures in 2100 eventuate (Wernberg et al., 2011; Martínez et al., 2018). In contrast, Arctic kelp forests of *Laminaria digitata* have extended with ocean

warming into areas that were previously too-cold (Bartsch et al., 2016). However, increases in UV radiation, sediment loading and freshwater inputs can also negatively impact the distribution of the Arctic kelp communities (Filbee-Dexter et al., 2019). Modelling of the future kelp biome distribution is necessary to predict its responses to climate change.

The kelp biome map indicates where kelp forests could occur. If kelp is absent, it may be due to ocean warming, high turbidity, and/or over-grazing following 'trophic cascades' caused by hunting and fishing of animals that predate sea urchins, and consequent over-grazing of kelp by the sea urchins (e.g., Leleu et al., 2012; Filbee-Dexter and Scheibling, 2014). Future studies may model the potential distribution of individual kelp species at local and regional scales, so as to provide finer spatial resolution for local scale conservation and fishery management. The availability of additional data for the species with insufficient data may allow their range to be mapped. The present map may also be a useful data layer for predicting the occurrence of kelp-associated species and estimating 'blue carbon' budgets. Moreover, knowing the global extent of the kelp biome is important for the conservation of not only kelp but associated species, including species threatened with extinction or important to fisheries. Thus, the present kelp biome area was included as one of the biodiversity layers in designing a global network of Marine Protected Areas (Zhao et al., 2020).

CRedit authorship contribution statement

Dinusha Rasanjalee Menike Jayatilake: Conceptualization, data collection, methodology, software, and writing.

Mark John Costello: Conceptualization, methodology, supervision, reviewing and editing.

Declaration of competing interest

The authors whose names are listed immediately below certify that they have NO affiliations with or involvement in any organization or entity with any financial interest (such as honoraria; educational grants; participation in speakers' bureaus; membership, employment, consultancies, stock ownership, or other equity interest; and expert testimony or patent-licensing arrangements), or non-financial interest (such as personal or professional relationships, affiliations, knowledge or beliefs) in the subject matter or materials discussed in this manuscript.

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Appendix A. Supplementary data

Supplementary data to this article can be found online at <https://doi.org/10.1016/j.biocon.2020.108815>.

References

- Alongi, D., 2018. Kelp forests. In: Blue Carbon. Springer Briefs in Climate Studies. Springer, Cham. https://doi.org/10.1007/978-3-319-91698-9_5.
- Assis, J., Lucas, A.V., Barbara, I., Serrão, E.A., 2016. Future climate change is predicted to shift long-term persistence zones in the cold-temperate kelp *Laminaria hyperborea*. Mar. Environ. Res. 113, 174–182. <https://doi.org/10.1016/j.marenvres.2015.11.005>.
- Bartsch, I., Paar, M., Fredriksen, S., et al., 2016. Changes in kelp forest biomass and depth distribution in Kongsfjorden, Svalbard, between 1996–1998 and 2012–2014 reflect Arctic warming. Polar Biol. 39, 2021–2036.
- Basher, Z., Costello, M.J., 2019. World maps of ocean environment variables. In: Encyclopedia of the World's Biomes. Reference Module in Earth Systems and Environmental Sciences. Elsevier, 11 pp. ISBN 9780124095489. <https://doi.org/10.1016/B978-0-12-409548-9.12076-7>.

- Basher, Z., Bowden, D.A., Costello, M.J., 2014a. Diversity and distribution of deep-sea shrimps in the Ross Sea region of Antarctica. *PLoS One* 9. <https://doi.org/10.1371/journal.pone.0103195>.
- Basher, Z., Bowden, D.A., Costello, M.J., 2014b. Global marine environment datasets (GMED). world wide web electronic publication. Version 1.0 (Rev.01.2014). accessed at <http://gmed.auckland.ac.nz>.
- Bo, M., Bertolino, M., Borghini, M., et al., 2011. Characteristics of the mesophotic megabenthic assemblages of the Vercelli Seamount (North Tyrrhenian Sea). *PLoS One* 6, 1–11.
- Costello, M.J., Zhao, Q., Jayatilake, D.R.M., 2020. Defining marine spatial units: realms, biomes, ecosystems, seascapes, habitats, biotopes, communities and guilds. In: *Encyclopedia of the World's Biomes. Reference Module in Earth Systems and Environmental Sciences*. Elsevier, 9 pp. ISBN 9780124095489. <https://doi.org/10.1016/B978-0-12-409548-9.12515-1>.
- Dayton, P.K., 1985. Ecology of kelp communities. *Annu. Rev. Ecol. Syst.* 16, 215–245. <https://doi.org/10.1146/annurev.es.16.110185.001243>.
- Duarte, C.M., 2017. Reviews and syntheses: hidden forests, the role of vegetated coastal habitats in the ocean carbon budget. *Biogeosciences* 14, 301–310. <https://doi.org/10.5194/bg-14-301-2017>.
- Elith, J., Graham, C.H., Anderson, R.P., Dudík, M., et al., 2006. Novel methods improve prediction of species' distributions from occurrence data. *Ecography* 29, 129–151.
- Elith, J., Phillips, S.J., Hastie, T., et al., 2011. A statistical explanation of MaxEnt for ecologists. *Divers. Distrib.* 17, 43–57.
- Espriella, M., Schaper, T., Atchia, A., Rose, K., Lecours, V., 2019. Habitat mapping of giant kelp (*Macrocystis pyrifera*) and devil weed (*Sargassum horneri*) off the coast of Santa Catalina Island, California. *McGill Sci. Undergrad. Res. J.* 14 (1).
- Filbee-Dexter, K., Scheibling, R.E., 2014. Sea urchin barrens as alternative stable states of collapsed kelp ecosystems. *Mar. Ecol. Prog. Ser.* 495, 1–25.
- Filbee-Dexter, K., Wernberg, T., 2018. Rise of turfs: a new battlefield for globally declining kelp forests. *BioScience* 68, 64–76.
- Filbee-Dexter, K., Wernberg, T., Fredriksen, S., Norderhaug, K.M., Pedersen, M.F., 2019. Arctic kelp forests: diversity, resilience and future. *Glob. Planet. Chang.* 172, 1–14.
- Fraser, C.I., Spencer, H.G., Waters, J.M., 2009. Glacial oceanographic contrasts explain phylogeography of Australian kelp. *Mol. Ecol.* 18, 2287–2296.
- Frid, C.L.J., Kitching, J.A., 1988. The *Laminaria* forest of Barloge Creek, Lough Hyne, Ireland, with special reference to the importance of wave action. *Irish Nat. J.* 22, 463–469.
- Gaylord, B., Reed, D.C., Raimondi, P.T., 2002. A physically based model of macroalgal spore dispersal in the wave and current-dominated nearshore. *Ecology* 83, 1239–1251. <https://doi.org/10.1890/00129658>.
- GBIF, 2017. Global biodiversity information facility occurrence download of Laminariales. <https://doi.org/10.15468/dl.oanp7d>. (Accessed 24 October 2017).
- Gerard, V.A., 1997. The role of nitrogen nutrition in high-temperature tolerance of kelp, *Laminaria saccharina* (Chromophyta). *J. Phycol.* 33, 800–810. <https://doi.org/10.1111/j.0022-3646.1997.00800.x>.
- Giri, C., Ochieng, E., Tieszen, L.L., et al., 2011. Status and distribution of mangrove forests of the world using earth observation satellite data (version 1.3, updated by UNEP-WCMC). *Glob. Ecol. Biogeogr.* 20, 154–159. <https://doi.org/10.1111/j.1466-8238.2010.00584.x>. Data URL: <http://data.unep-wcmc.org/datasets/4>. (Accessed 7 June 2016).
- Graham, M.H., Harrold, C., Lysin, S., et al., 1997. Population dynamics of giant kelp *Macrocystis pyrifera* along a wave exposure gradient. *Mar. Ecol. Prog. Ser.* 148, 269–279. <https://doi.org/10.3354/meps148269>.
- Graham, M.H., Kinlan, B.P., Druehl, L.D., et al., 2007. Deep-water kelp refugia as potential hotspots of tropical marine diversity and productivity. *Proc. Natl. Acad. Sci. U. S. A.* 104, 16576–16580. <https://doi.org/10.1073/pnas.0704778104>.
- Guiry, M.D., Guiry, G.M., 2018. *AlgaeBase*. World-wide Electronic Publication. National University of Ireland, Galway. <http://www.algaebase.org>. (Accessed 11 October 2018).
- Guiry, M.D., Guiry, G.M., 2020. *AlgaeBase*. World-wide Electronic Publication. National University of Ireland, Galway. <http://www.algaebase.org>. (Accessed 12 June 2020).
- Horton, T., Kroh, A., Ah Yong, S., Bailly, N., Boyko, C.B., et al., 2020. World register of marine species. Available at VLIZ from <https://www.marinespecies.org>. (Accessed 21 April 2020).
- Hurd, C.L., 2000. Water motion, marine macroalgal physiology, and production. *J. Phycol.* 36, 453–472. <https://doi.org/10.1046/j.1529-8817.2000.99139.x>.
- Jayatilake, D.R.M., Costello, M.J., 2018. A modelled global distribution of the seagrass biome. *Biol. Conserv.* 226, 120–126. <https://doi.org/10.1016/j.biocon.2018.07.009>.
- Jayatilake, D.R.M., Costello, M.J., 2020a. Model validation data from 2018 to 2020. In: Dataset. Figshare. The University of Auckland. <https://doi.org/10.17608/k6.auckland.12278786.v2>.
- Jayatilake, D.R.M., Costello, M.J., 2020b. Training occurrence records of the MaxEnt model. In: Dataset. Figshare. The University of Auckland. <https://doi.org/10.17608/k6.auckland.12272033.v1>.
- Jayatilake, D.R.M., Costello, M.J., 2020c. The kelp biome. In: *Encyclopedia of the World's Biomes*. <https://doi.org/10.1016/B978-0-12-409548-9.11768-3>.
- Kraan, S., Tramullas, A.V., Guiry, M.D., 2000. The edible brown seaweed *Alaria esculenta* (Phaeophyceae, Laminariales): hybridization, growth and genetic comparisons of six Irish populations. *J. Appl. Phycol.* 12, 577–583.
- Krause-Jensen, D., Duarte, C.M., 2016. Substantial role of macroalgae in marine carbon sequestration. *Nat. Geosci.* 9, 737–742.
- Krumhansl, K.A., Scheibling, R.E., 2012. Production and fate of kelp detritus. *Mar. Ecol. Prog. Ser.* 467, 281–302.
- Krumhansl, K.A., Okamoto, D.K., Rassweiler, A., et al., 2016. Global patterns of kelp forest change over the past half-century. *Proc. Natl. Acad. Sci. U. S. A.* 113, 13785–13790. <https://doi.org/10.1073/pnas.1606102113>.
- Leleu, K., Remy-Zephir, B., Grace, R., Costello, M.J., 2012. Mapping habitat change after 30 years in a marine reserve shows how fishing can alter ecosystem structure. *Biol. Conserv.* 155, 193–201.
- Lüning, K., 1990. *Seaweeds. In: Their Environment, Biogeography and Ecophysiology*. John Wiley & Sons, Inc., New York.
- Mann, K.H., 1973. Seaweeds: their productivity and strategy for growth. *Science* 182, 975–981.
- Marins, B.V., Amado-Filho, G.M., Barreto, M.B., Longo, L.L., 2012. Taxonomy of the southwestern Atlantic endemic kelp: *Laminaria abyssalis* and *Laminaria brasiliensis* (Phaeophyceae, Laminariales) are not different species. *Phycol. Res.* 60 (1), 51–60.
- Markel, R.W., Shurin, J.B., 2015. Indirect effects of sea otters on rockfish (*Sebastes* spp.) in giant kelp forests. *Ecology* 96, 2877–2890.
- Martínez, B., Wernberg, T., Radford, B., Thomsen, M.S., et al., 2018. Distribution models predict large contractions of habitat-forming seaweeds in response to ocean warming. *Divers. Distrib.* 24, 1350–1366.
- Matsunaga, K., Kawaguchi, T., Suzuki, Y., Nigi, G., 1999. The role of terrestrial humic substances on the shift of kelp community to crustose coralline algae community of the southern Hokkaido Island in the Japan Sea. *J. Exp. Mar. Biol. Ecol.* 241, 193–205.
- Moe, R.L., Silva, P.C., 1977. Antarctic marine flora: uniquely devoid of kelps. *Science* 196, 1206–1208.
- Muller, R., Laepple, T., Bartsch, I., Wiencke, C., 2009. Impact of oceanic warming on the distribution of seaweeds in polar and cold-temperate waters. *Bot. Mar.* 52, 617–638. <https://doi.org/10.1515/BOT.2009.080>.
- Norton, T.A., 1992. Dispersal by macroalgae. *Br. Phycol. J.* 27 (3), 293–301. <https://doi.org/10.1080/00071619200650271>.
- OBIS, 2017. Ocean Biogeographic Information System. Occurrence download of order Laminariales. Retrieved from: <http://www.iobis.org>. (Accessed 24 October 2017).
- Parker, T., Tunnicliffe, V., 1994. Dispersal strategies of the biota on an oceanic seamount: implications for ecology and biogeography. *Biol. Bull.* 187, 336–345.
- Paula, B.M., Marcomini, S.C., Casas, G.N., 2018. Environmental impacts of an alien kelp species (*Undaria pinnatifida*, Laminariales) along the Patagonian Coasts. In: Makowski, C., Finkl, C. (Eds.), *Impacts of Invasive Species on Coastal Environments*. Coastal Research Library, vol. 29. Springer, Cham. https://doi.org/10.1007/978-3-319-91382-7_10.
- Pedersen, M.F., Nejrup, L.B., Fredriksen, S., et al., 2012. Effects of wave exposure on population structure, demography, biomass and productivity of the kelp *Laminaria hyperborea*. *Mar. Ecol. Prog. Ser.* 451, 45–60.
- Pereyra, P.J., de la Barra, P., Gastaldi, M., et al., 2017. When the tiny help the mighty: facilitation between two introduced species, a solitary ascidian and a macroalga in northern Patagonia, Argentina. *Mar. Biol.* 164–185. <https://doi.org/10.1007/s00227-017-3202-1>.
- Peteiro, C., Freire, Ó., 2012. Outplanting time and methodologies related to mariculture of the edible kelp *Undaria pinnatifida* in the Atlantic coast of Spain. *J. Appl. Phycol.* 24 (6), 1361–1372.
- Peterson, A.T., Papes, M., Soberon, J., 2008. Rethinking receiver operating characteristic analysis application in ecological niche modelling. *Ecol. Model.* 213, 63–72.
- Peterson, A.T., Soberon, J., Pearson, R.G., et al., 2011. *Ecological Niches and Geographic Distributions*. Princeton University Press, United States of America.
- Phillips, S.J., 2017. A brief tutorial on Maxent. AT&T Labs-Research. Available from url: http://biodiversityinformatics.amnh.org/open_source/maxent. (Accessed 10 January 2019).
- Phillips, S.J., Dudík, M., 2008. Modeling of species distributions with Maxent: new extensions and a comprehensive evaluation. *Ecography* 31, 161–175. <https://doi.org/10.1111/j.0906-7590.2008.5203.x>.
- Phillips, S.J., Dudík, M., Schapire, R.E., 2004. A maximum entropy approach to species distribution modeling. In: *Proceedings of the 21st International Conference on Machine Learning*. ACM Press, New York, pp. 655–662.
- Phillips, S.J., Anderson, R.P., Schapire, R.E., 2006. Maximum entropy modeling of species geographic distributions. *Ecol. Model.* 190, 231–259. <https://doi.org/10.1016/j.ecolmodel.2005.03.026>.
- Quartino, M.L., Boraso de Zaixso, A.L., 2008. Summer macroalgal biomass in Potter Cove, South Shetland Islands, Antarctica: its production and flux to the ecosystem. *Polar Biol.* 31, 281–294.
- Raffaelli, D., Hawkins, S., 1999. *Intertidal Ecology*. Kluwer Academic Publishers, The Netherlands.
- Raffo, M.P., Eyrales, M.C., Iribarne, O.O., 2009. The invasion of *Undaria pinnatifida* to a *Macrocystis pyrifera* kelp in Patagonia (Argentina, south-west Atlantic). *J. Mar. Biol. Assoc. U. K.* 89, 1571–1580. <https://doi.org/10.1017/S002531540900071X>.
- Raybaud, V., Beaugrand, G., Goberville, E., et al., 2013. Decline in kelp in West Europe and climate. *PLoS One* 8 (6), e66044.
- Saeedi, H., Dennis, T.E., Costello, M.J., 2016. Bimodal latitudinal species richness and high endemism of razor clams (Mollusca). *J. Biogeogr.* 44, 592604. <https://doi.org/10.1111/jbi.12903>.
- Shao, Z., Wang, W., Zhang, P., Yao, J., Wang, F., et al., 2019. Genome-wide identification of genes involved in carbon fixation in *Saccharina japonica* and responses of putative C4-related genes to bicarbonate concentration and light intensity. *Plant Physiol. Biochem.* 137, 75–83.
- Smale, D.A., Burrows, M.T., Moore, P., O'Connor, N., Hawkins, S.J., 2013. Threats and knowledge gaps for ecosystem services provided by kelp forests: a northeast Atlantic perspective. *Ecol. Evol.* 3, 4016–4038. <https://doi.org/10.1002/ece3.774>.
- Starko, S., Gomez, M.S., Darby, H., et al., 2019. A comprehensive kelp phylogeny sheds light on the evolution of an ecosystem. *Mol. Phylogenet. Evol.* 136, 138–150.
- Steneck, R.S., Graham, M.H., Bourque, B.J., et al., 2002. Kelp forest ecosystems: biodiversity, stability, resilience and future. *Environ. Conserv.* 29, 436–459. <https://doi.org/10.1017/S0376892902000322>.

- Stévant, P., Marfaing, H., Duinker, A., et al., 2018. Biomass soaking treatments to reduce potentially undesirable compounds in the edible seaweeds sugar kelp (*Saccharina latissima*) and winged kelp (*Alaria esculenta*) and health risk estimation for human consumption. *J. Appl. Phycol.* 30 (3), 2047–2060.
- Teagle, H., Hawkins, S.J., Moore, P.J., Smale, D.A., 2017. The role of kelp species as biogenic habitat formers in coastal marine ecosystems. *J. Exp. Mar. Biol. Ecol.* 492, 81–98. <https://doi.org/10.1016/j.jembe.2017.01.017>.
- Tittensor, D.P., Baco, A.R., Brewin, P.E., et al., 2009. Predicting global habitat suitability for stony corals on seamounts. *J. Biogeogr.* 36, 1111–1128. <https://doi.org/10.1111/j.1365-2699.2008.02062.x>.
- UNEP-WCMC, WorldFish Centre, WRI, TNC, 2018. Global Distribution of Warm-Water Coral Reefs, Compiled From Multiple Sources Including the Millennium Coral Reef Mapping Project. Version 4.0. Includes Contributions From IMaRS-USF and IRD (2005), IMaRS-USF (2005) and Spalding et al. (2001). UN Environment World Conservation Monitoring Centre, Cambridge (UK). URL: <http://data.unep-wcmc.org/datasets/1>. (Accessed 24 April 2019).
- Vásquez, J.A., Zuñiga, S., Tala, F., et al., 2014. Economic valuation of kelp forests in northern Chile: values of goods and services of the ecosystem. *J. Appl. Phycol.* 26, 1081–1088. <https://doi.org/10.1007/s10811-013-0173-6>.
- Verbruggen, H., Tyberghein, L., Pauly, K., et al., 2009. Macroecology meets macroevolution: evolutionary niche dynamics in the seaweed *Halimeda*. *Glob. Ecol. Biogeogr.* 18, 393–405. <https://doi.org/10.1111/j.1466-8238.2009.00463.x>.
- Wernberg, T., Filbee-Dexter, K., 2019. Missing the marine forest for the trees. *Mar. Ecol. Prog. Ser.* 612, 209–215. <https://doi.org/10.3354/meps12867>.
- Wernberg, T., Thomsen, M.S., 2005. The effect of wave exposure on the morphology of *Ecklonia radiata*. *Aquat. Bot.* 83, 61–70.
- Wernberg, T., Russell, B.D., Moore, P.J., et al., 2011. Impacts of climate change in a global hotspot for temperate marine biodiversity and ocean warming. *J. Exp. Mar. Biol. Ecol.* 400, 7–16. <https://doi.org/10.1016/j.jembe.2011.02.021>.
- Wernberg, T., Smale, D.A., Tuya, F., et al., 2012. An extreme climatic event alters marine ecosystem structure in a global biodiversity hotspot. *Nat. Clim. Chang.* 3, 78–82.
- Wernberg, T., Bennett, S., Babcock, R.C., et al., 2016. Climate-driven regime shift of a temperate marine ecosystem. *Science* 353, 169–172. <https://doi.org/10.1126/science.aad8745>.
- Wernberg, T., Krumhansl, K.A., Filbee-Dexter, K., Pedersen, M.F., 2019. Status and trends for the world's kelp forests. In: Sheppard, C. (Ed.), *World Seas: An Environmental Evaluation*, Second edition vol. III. Elsevier. <https://doi.org/10.1016/B978-0-12-805052-1.00003-6>.
- Włodarska-Kowalczyk, M., Kukliński, P., Ronowicz, M., et al., 2009. Assessing species richness of macrofauna associated with macroalgae in Arctic kelp forests (Hornsund, Svalbard). *Polar Biol.* 32, 897–905. <https://doi.org/10.1007/s00300-009-0590-9>.
- Woodward, F.I., Lomas, M.R., Kelly, C.K., 2004. Global climate and the distribution of plant biomes. *Philos. Trans. R. Soc. Lond. B Biol. Sci.* 359 (1450), 1465–1476. <https://doi.org/10.1098/rstb.2004.1525>.
- Wu, Z., Zhang, X., Lozano-Montes, H.M., Loneragan, N.R., 2016. Trophic flows, kelp culture and fisheries in the marine ecosystem of an artificial reef zone in the Yellow Sea. *Estuar. Coast. Shelf Sci.* 182, 86–97.
- Yesson, C., Taylor, M.L., Tittensor, D.P., et al., 2012. Global habitat suitability of cold-water octocorals. *J. Biogeogr.* 39, 1278–1292. <https://doi.org/10.1111/j.1365-2699.2011.02681.x>.
- Zhao, Q., Stephenson, F., Lundquist, C., Kaschner, K., Jayatilake, D.R.M., Costello, M.J., 2020. Where marine protected areas would best represent 30% of ocean biodiversity. *Biol. Conserv.* 244, 108536. <https://doi.org/10.1016/j.biocon.2020.108536>.
- Žuljević, A., Peters, A.F., Nikolić, V., et al., 2016. The Mediterranean deep-water kelp *Laminaria rodriguezii* is an endangered species in the Adriatic Sea. *Mar. Biol.* 163, 69. <https://doi.org/10.1007/s00227-016-2821-2>.